

# POSSIBILITIES FOR CONTROLLING THE CONNECTION BETWEEN WASTEWATER TREATMENT PLANTS AND RECEIVING WATER BODIES

J. A. ROSERO and K. BUZÁS

Department of Hydraulic Engineering,  
Technical University, H-1521 Budapest

Received July 14, 1986

Presented by Prof. Dr. M. Kozák

## Abstract

With increasing water supply and canalization the quality of the water playing the role of receiver often deteriorates. Among other problems, the not sufficient removal of plant nutrients by the traditional purification technology causes difficulties. Considering that the removal of fertilizers results in a significant increase in the costs of wastewater treatment, it is expedient to keep the extent of removal just at the yet acceptable level of the pollutant.

This paper deals with the removal of phosphorus as fertilizer on a transport theoretical basis.

As a result, a qualitative connection is established between water quality and water management on the one hand and the operation of the wastewater treatment plant on the other hand.

## The objectives and method of the research work

With the increase in water supply and canalization the quality of the receiving water bodies often deteriorates, not only in Hungary, but all over the world. This deterioration is caused at present mainly by the introduction of the wastewater of dwellings into the receiving water bodies in spite of the increasing number of wastewater treatment plants being built. According to the literature and experience, the main reason of this problem is that — in addition to other effects as e.g. the not point-like nature of pollutants — traditional purification procedures do not remove fertilizers satisfactorily. The enrichment of these substances, in turn, may lead to the breaking of the ecological equilibrium in the receiving water body. Since the removal of fertilizers results in a rapid increase in the costs of purification, it is expedient to keep the level of pollutants at the allowed value. The necessary extent of removal can be determined from the receiving capacity for the fertilizer in question. Therefore it is justified technically and economically to analyse the ecology of the receiver and the procedure of wastewater treatment as well as the connection between them in order to make a sound foundation of design and operation.

As it became clear from earlier, mainly from biological research, the main energy source of eutrophication processes is phosphorus, problems of the transport of phosphorus have been investigated in our study. Upon considering the economically feasible solutions here and in other countries, emphasis was

given to the technical realization of results achievable by completing the purification processes utilizing activated sludge.

As a general goal, the description of the purification technology of communal wastewater by activated sludge (AS) has been set. A parametric treatment is elaborated for the third purification step with special regard to the removal of fertilizers ( $P$ ), by starting from observations concerning the effect of the dynamic-kinetic change of AS on the AS base line.

The normalization of the system may increase the efficiency of the secondary clarifier when considering the macro- and submacro-arrangements and the microdynamics of differentiated sedimentation. It also improves the relation of the wastewater treatment plant and the receiving water body, since at the separation of phases the mechanism of mass transport during adsorption-storage and the kinetics of coagulation-flocculation may be brought in harmony by the consideration of the normal distributions of critical concentrations and time.

Other components of the connection between the wastewater treatment plant and the receiving water body, as well as the composite processes of technology, such as e.g. filtration, ozonization desinfection, etc. will be neglected here.

Field experiments were carried out in the wastewater treatment plant at Keszthely based on the critical evaluation of technical literature. The analysis and evaluation of the experiments occurred by the use of the ITO (Input, Transformation, Output) system. Technical literature was considered from the viewpoint of transport theory.

### System approach in the removal of phosphorus

The results of field experiments carried out in 1983—84 were compared with the data found in the literature. Correlations thus obtained were classified by using transport theory.

The overall equation system of the conditions for the removal of phosphorus at macro-, micro- and submicrolevels of purification processes using activated sludge is the following:

$$C_f = F [e | i_f ] \quad (1)$$

$$K_{C_a} = C_f | t | \quad (2)$$

$$C_f | t | = F [Y_{X_k}] \quad (3)$$

$$Y_h = F [X_k | AS_{p,p'}] \quad (4)$$

Let the transformation interpreting the removal of phosphorus be

$$X_k = F | AS_{p,p'} |$$

what is a correlation between the mass of flocs and that of the wastewater in its environment (Fig. 1).

The removal of phosphorus may occur by either biological assimilation or by the combination of this and chemical precipitation, the time course of which may be characterized by distribution factor  $K_{C_z}$  together with equations

$$e | i' p | = \left[ \gamma \frac{K_{OI}}{P} ; \beta \frac{BOI}{P} \right] \rightarrow \max \tag{5}$$

$$i''' = \left[ \alpha \frac{K_{OI}}{P} ; \delta \frac{BOI}{P} \right] \rightarrow \max \tag{6}$$

where each

$$i'''_p = [i_j \text{ Poly-P} \leq i_k \text{ ATP}] \tag{7}$$

$$i'''_p = [pH; i_{cat}; i_{inh}], \text{ in which} \tag{8}$$

$i_{cat}$  is the intensive of cations and

$$i_{cat} = i [Ca, Fe, Al] \rightarrow \max \tag{9}$$

$i_{inh}$  = the intensive of inhibitors

$$i_{inh} = i [Mg^{2+}, \text{ bicarbonate, Poly-P}] \rightarrow \min \tag{10}$$

The storage of Poly-P is a precondition for choosing arrangement  $X_k$  with respect to  $l/T$  which ensures the variation of the anaerobic-aerobic states of the sludge.

The effect of this change in  $X_k$  appears as a variation in intensive  $i_p$  at the submicro level according to

$$X_k = F(AS_{p,p'})$$

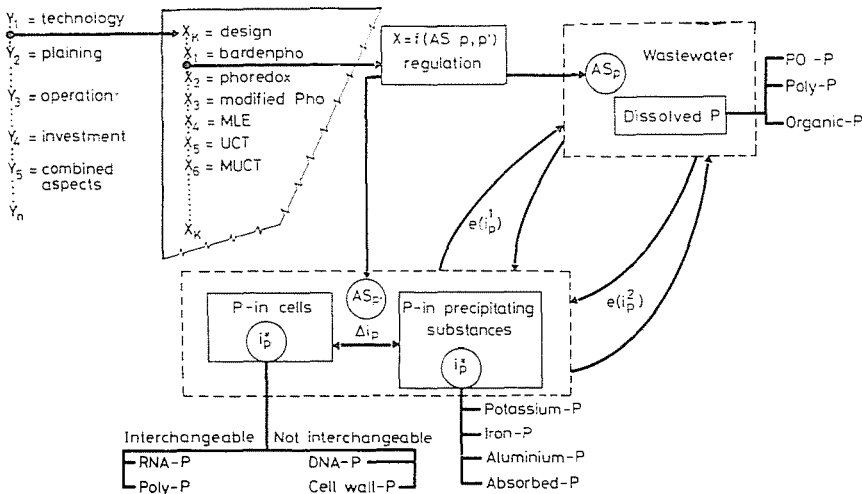


Fig. 1. System approach in the removal of phosphorus

The acceleration or deceleration of

$$i_p = [i_p^1; i_p^2]$$

shows the effect of the introduction of arrangement  $X_k$  in the distribution of  $K_{C_z} P$  (Fig. 2). From this it can be concluded that in aerobic systems where the transport of energy and matter occurs according to the equation

$$AS_p[e | i_p^2], AS_{p'}[e | i_p^1] \rightarrow K_{C_z} \quad (11)$$

phosphorus is also precipitated (Figs 1, 2). Thus  $K_{C_z}$  could be measured experimentally what makes the solution of equation system (1)–(4) possible.

Thus from the viewpoint of the transport of energy and matter, five mechanisms can be set as the most probable ways of the biologically assisted chemical removal of phosphorus:

1. The basic mechanism is the normal assimilation of phosphorus which is always operating and the efficiency of which depends on  $i_p$  (it results 1–2,5% in an  $e | i_p^1$ ).

2. The storage of Poly-P which is ensured by changing the anaerobic and aerobic environments of the activated sludge, in dependence of  $i_p$  (A storage of  $P$  as high as 5–8% may be achieved).

3. Ordinary precipitation in the wastewater occurring due to the change in the aerobic-anoxic-aerobic states as a function of  $i_p''$ .

4. Accelerated precipitation in the wastewater when function  $i_p''$  is near to its maximum value or even reaches it.

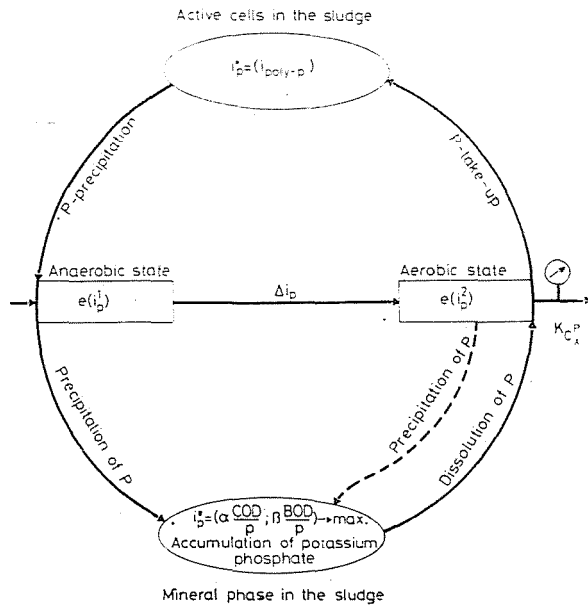


Fig. 2. Scheme of the system intensively induced for P

Biofilm precipitation occurring in the case of bacterial denitrification if the wastewater environment of the biofilm is favourable, in dependence of  $i_p''$ .

The bonity of the optimization of functions and the aspects of the given system depend strongly on the correctness and consistence of qualitative and quantitative data required in the application of the different mechanisms.

**The role of the differentiated settling in the transport of nutrients**

Since the main energy source of eutrophisation is phosphorus, problems concerning the transport of phosphorus were studied.

The connection between the wastewater treatment plant and the receiving water body is discussed on the basis of the ecology of the above mentioned nutrient-transport and the system of the purification system with activated sludge which was examined from the viewpoint of the phosphorus removal (ITO/AS<sub>C</sub>). The scheme of the system is illustrated in Fig. 3.

The most sensitive unit of the system is subsystem 3. This is partly due to the strongly stochastic characteristics of the input intensives,  $T_p$  and  $T_C$  in transport  $T_{23}$ . The intensives of the AS base line are partly also quasi-stochastic.

In the general case, the output of subsystem 3 is the output of the secondary clarifier.

The observation of this change in the transition section of the settling curve (Fig. 4) and its mathematical modelling makes possible the observation

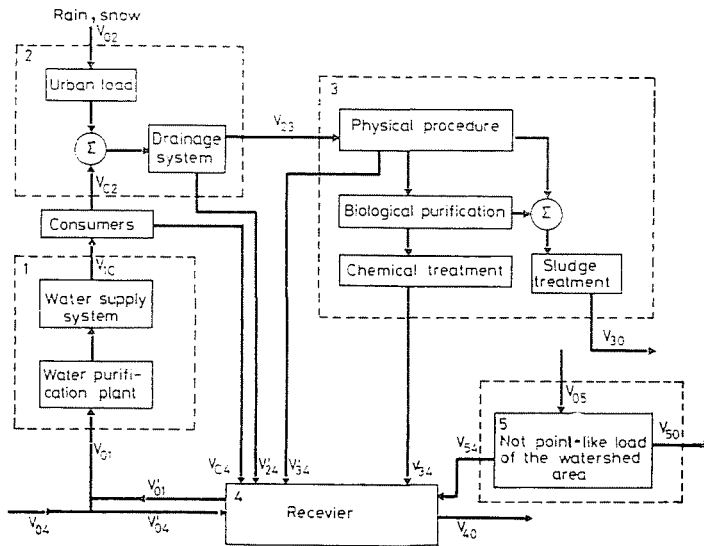


Fig. 3. Connection between the wastewater treatment plant and the receiver

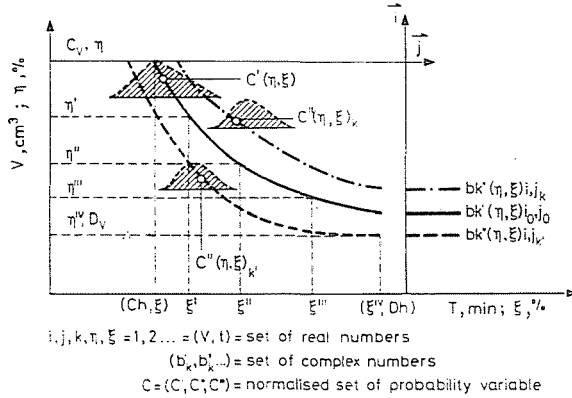


Fig. 4. Mathematical modelling of the transition reaction

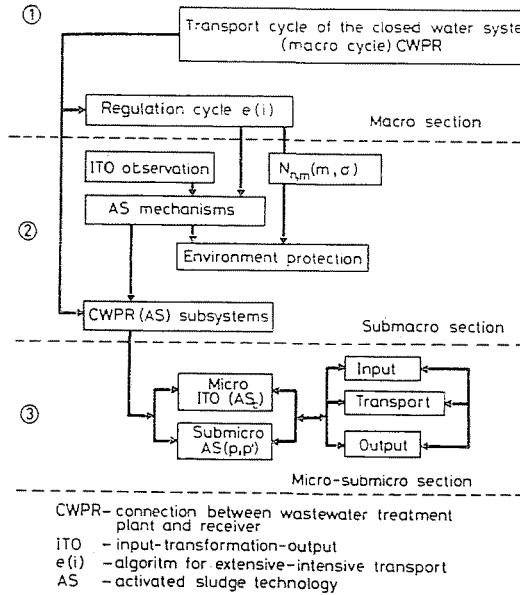


Fig. 5

of  $AS_C$  from the viewpoint of phase separation. In turn, this provides a possibility for the regulation of the connection between the wastewater treatment plant and the receiving water body by means of intensive  $k_c$ .

Disregarding the effect of intensive  $V_{30}$  on the environment, this study is aimed at investigating the changing process of changing of matter-energy intensive  $V_{34}$ .

The incorporation of the mathematical model of settling process into the connection system wastewater treatment plant — receiving water body oc-

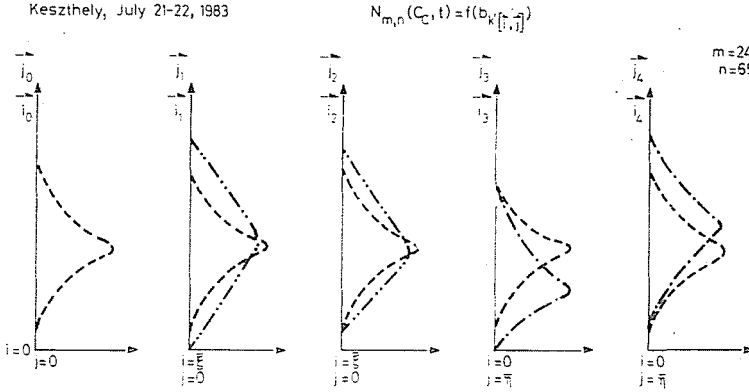


Fig. 6. Effect of (i, j) system on (system)

currred by the utilization of the current biochemical results from the study of purification processes with activated sludge. This was the basis for designing our system. This system applies the submicro, micro, submacro and macro sections for describing the operation of ITO/AS<sub>c</sub> (Fig. 5).

One of the possible approaches for characterizing ITO/AS<sub>c</sub> is the construction of the interactions of the (i, j), (ξ, η) and N(C<sub>c</sub>, t<sub>c</sub>) systems. System (i, j) forms space, (ξ, η) represents the description of the stochastic flow of matter and energy in time, and regulating system N(C<sub>c</sub>, t<sub>c</sub>) ensures the possibility of standardization of the whole ITO/AS system.

The result of this interaction is the distribution k<sub>c</sub> (Fig. 6) which shows that differentiated settling manifests itself in different effect-differences in the directions of the two axes. The existence construction and utilization of this distribution occurs in inductive and deductive ways based on experience from experiments and data from the literature. By using this distribution

$$(Q_{ke} - Q_{kc}) \text{ and } (V_{p \min} - A_{p \max})$$

can be given for the separation of phases, whereas the resultant differentiated settling rate may be written as

$$v_p = K_c \sqrt{\frac{2g}{k_c C_D}} \text{ (m/s)} \tag{12}$$

where

$$C_D = \frac{1}{C_D P^2 ; P} \text{ (cm}^2\text{/g), as } C_D \text{ is an}$$

inverse function of mass flow from the viewpoint of differentiated settling. C<sub>Dp<sup>2</sup>;p</sub> = roughness factor in which the characteristics of submacro system

DS<sub>ke</sub> - CF<sub>ke</sub> are reflected,  
 DS = adaptive-learning mechanism

$CF$  = coagulation-flocculation mechanism  
 $k_e$  = effect of the above mechanism on ailing  
 $k_c$  = effect of the above mechanism on secondary clarifier

$$C_{D_p; p} = \begin{bmatrix} f | DS_{ke} ; CF_{kc} | \\ f | DS_{ke} ; CF_{ke} | \end{bmatrix}$$

Thus the equilibrium conditions of the initial state can be written in the following new form:

$$F_D = f v_p (C_D) ; A_{pmin} / v_{pmax} ; A S(|ke|) \quad (13)$$

The unknown  $C_D$  cannot be determined for individual flocculating particles, at least not in a deterministic way as the general curves ( $C_D$ , Re) are unknown for the third step of the AS technology. Hence if we consider that during the phase-separation process in the microsection the transport mechanism primarily characteristic for AS (the so-called adsorption-storage cycle) does not play a secondary role with respect to the coagulation-flocculation kinetic mechanism in the separation of phases as it is proved by photomicrographs, but it appears as a harmonic equilibrium alternative being the most sensitive indicator of the equilibrium state, then in the secondary clarifying process  $C_D$  would be

$$C_D = f(DS_{ke}, CF_{kc}) \quad (14)$$

and at simultaneous precipitation

$$C_D = f(DS_{ke}, CF_{ke}) \quad (15)$$

This is advantageous from the viewpoint of theoretical treatment, since intensive  $AS/k_c$  may be determined in an empirical way by direct measurements.

### Regulation possibilities of the connection between wastewater treatment plants and receiving water bodies

In addition to intensives ( $Q_{ke} - Q_{kc}$ ); ( $V_{p \min} - A_{p \max}$ ) appearing in phase separation, the normalising effect of  $ITO'AS_c/N(C_c, t_c)$  in  $(\xi, \eta)$  may be interpreted as a smoothing effect on the surface of flocculae due to the effect of different coagulating agents.

This "smoothing", erosion, and fraction, percussion and aggregation of the surface can be characterized by probabilities  $\vec{m}(\xi, \eta)$ . The resultant probability,  $\vec{M}(\xi, \eta)$  may be obtained by aggregating the probabilities of individual processes:

$$\vec{M}(\xi, \eta) = \lambda \vec{F}_R(\xi, \eta) = \Sigma F_R(\xi, \eta) \text{ and} \quad (16)$$

$$\vec{M}(\xi, \eta) = \Sigma \vec{m}(\xi, \eta) \quad (17)$$



In transport,  $V(23, 34)$ , vector  $M$  provides the total amount of work used for purification. The direction and numerical value of vector  $\vec{M}$  may be utilized in the synchronization of airing and secondary clarifier in subsystem 3 (Fig. 3), what means a possibility for improving the efficiency of the plant.

With vectors  $\vec{M}$  determined for plants operating with activated sludge, in larger regions a vector polygon can be constructed which provides information on the resultant movement of the AS systems in the region and makes the setting of regional ecological balance possible for the determination of the extent and direction of regulation required.

One of the means for regulating the connection between the wastewater treatment plant and the receiving water body is the distribution of load  $N_v(\eta, \sigma)$ .

In Figure 7 the distribution of the optimum intensive  $k_0$  for phase separation is shown in %. From this, the difference in loads can be forecasted for the mass transport  $V(OC, 2; 34)$ . The pointing out of mechanisms ensuring the dynamic equilibrium of the system became necessary for establishing just this.

1. We introduce the concept of delayed transport process  $T_d$  (Fig. 8). Then

$$V_{34}(t) = f(V_{02}(t - T_d); V_{C2}(t - T_d)) \tag{18}$$

where function  $f$  is defined in coordination system  $(i, j)$ , i.e. it is of extensive nature, correspondingly to Fig. 8A.

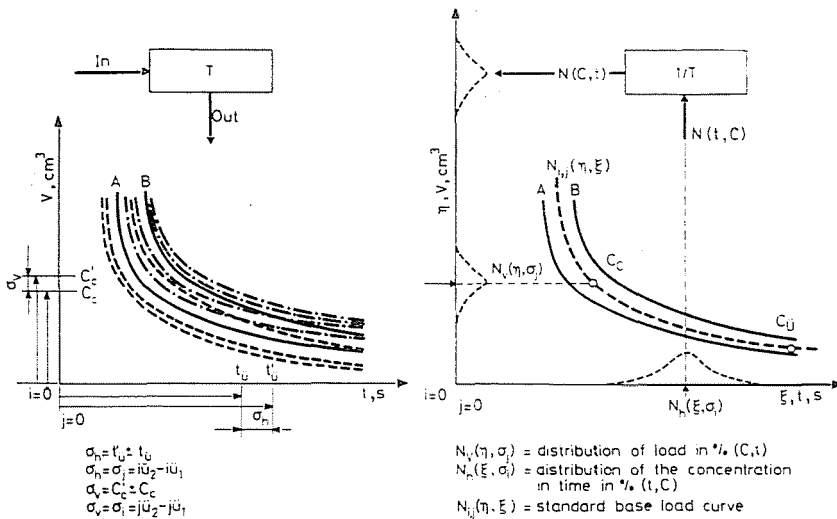


Fig. 7

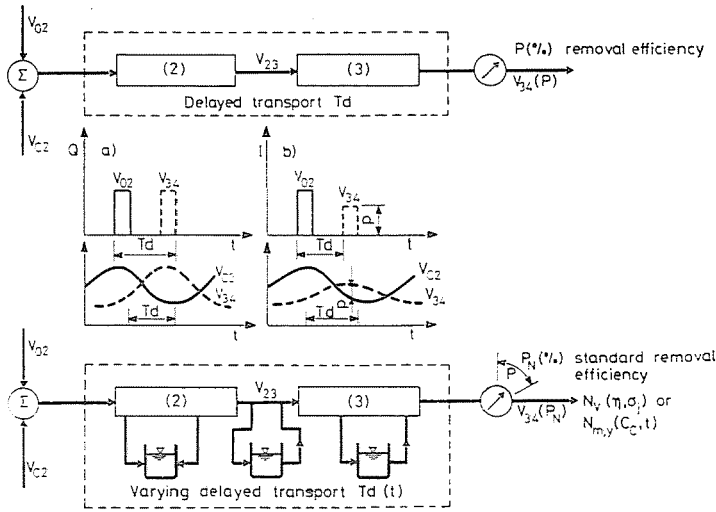
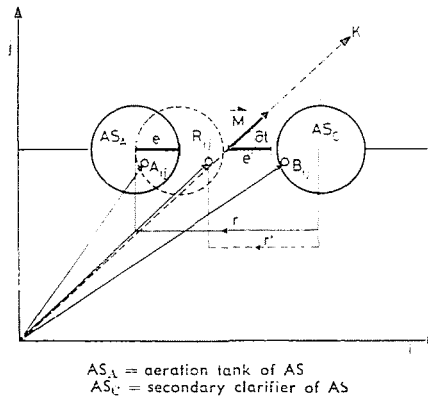


Fig. 8. Regulation of subsystems 2 and 3



AS<sub>x</sub> = aeration tank of AS  
 AS<sub>z</sub> = secondary clarifier of AS

Fig. 9. Subsystem of  $e =$  operation shift and  $r = f(e)$  recirculation in the macrosection of system (i, j)

2. In the macrosection of subsystem 3 the peaks in load different from standards can be smoothed by recirculation and simultaneous temporary storage as is shown in Fig. 9 This requires the adjustment of  $T_d$  according to the function of the intensive  $p$  (Fig. 8B), where  $p$  is the probability of the removal of loading pollutants. Optimum removal efficiency can be established obviously only when knowing the load  $V_{23}(p)$ .

In the first case the time of transport is invariant, since the distribution of output  $V_{34}(p)$  is similar to that of input  $V_{23}(p)$ . This is not consistent with the other parameters of the connection system of wastewater treatment plants

and receiving water body. Hence, considering that  $T_d(p)$  is not time-invariant, the load intensive of  $0 < p(\xi, \eta) < 100\%$  is given by distribution  $N_{m, n}(C_c, t)$  (Fig. 7), as the optimum efficiency of the plant in system  $(i, j)$  (Fig. 8).

The  $V_{23}(p) > N_{m, n}(C_c, t)$  means a state load which requires modification of  $T_d(p)$  by temporary storage or recirculation. The role of recirculation is illustrated in Fig. 9. Storage may be realized similarly to a facultative pool before subsystem 3 by  $T_d(t)_3$  and after it by  $T_d(t)_4$ . It may also be realized inside the subsystem on the basis of the standard removal efficiency  $P_N(\%)$  (Fig. 8).

Considering all this and based on earlier experience the connection between the wastewater treatment plant and the receiving water body, the analysis of the flow of matter  $m(V_{34}, V_{40})$  is written with the transport approximation

$$m[(i, e)_{34}; (i, e)_{40}] \quad (19)$$

### Interaction of water management and water quality in the connection of wastewater treatment plant and receiving water body

The description of the above connection has been attempted on the basis of the law of the conservation of matter by using the intensive parameters of mass transport. It is supposed that the yearly normal distribution of the intensive characterising the receiver water body can be given and let it be

$$N_m(i_{40}; e).$$

If the ratio of extensive is nearly constant in flow  $(V_{34}, V_{40})$ , the distribution of the intensives of the transport from the side of the plant is characterised by  $N_{m, n}(C_c, t) = N_m(i_{34}, e)$ .

Thus

$$N_m(\sigma_{i,e}) = N_m(i_{40}; e) - N_{m,n}(C_c, t) \quad (20)$$

In this case information can be obtained concerning the necessary intervention for synchronising the plant and the receiving water body in order to protect water quality.

The general form of Eq. (20) is:

$$N_m [i_{34}, e; i_{43}, e] \text{ where } e \approx k \quad (21)$$

If this is solved so that the extensives are considered state variants, then information can be obtained for water management. The general form of the equation may be given as:

$$N_m [e_{34}, i; e_{40}, i] \text{ where } i \approx k \quad (22)$$

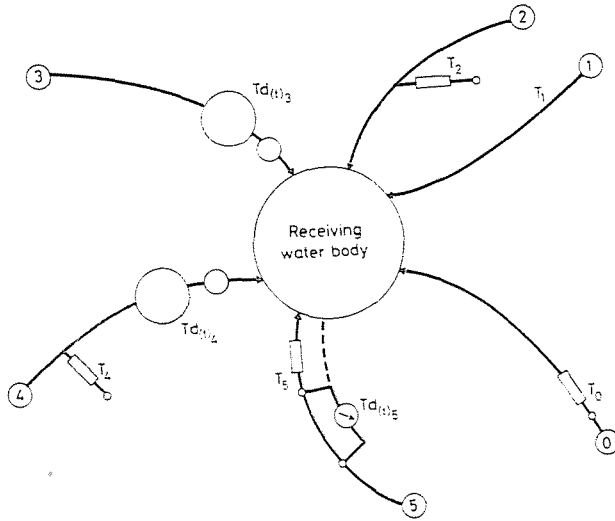


Fig. 10. Connection to the receiving system

From this it follows that the total solution of the connection provides the complete exploration of the interaction between water quality and water management Fig. 10. The most general form of this interaction can be written as

$$N_m[e | i |_{34} ; e | i |_{40}] \text{ where } e, i \neq k \tag{23}$$

The detailed elaboration of the practical applicability of this latter correlation is the most important task of future research.

### Denotations

- $c$  = change of the intensive
- $c_p$  = intensive change of phosphorus
- $AS_p$  = effect of the technological subsystem of  $AS$  macrosection on the liquid around the organic mass
- $AS_{p'}$  = effect of the technological subsystem of  $AS$  macrosection on the organic mass
- $e$  = amount of matter or energy
- $e(ip)$  = amount of matter or energy in function of the intensive of phosphorus
- $k$  = kinetic change in the transport of matter,- energy
- $k_c$  = time distribution of the kinetic change in intensive  $c$
- $Y_n$  =  $AS$  macrosection, vector of its subsystem
- $Y_1$  = technology

- $Y_2$  = planned  
 $Y_3$  = operational  
 $X_k$  = arrangement of  $k$  in the subsystem of  $Y$   
 $Y_1(x_1)$  = bardenpho  
 $Y_1(x_2)$  = Phoredox  
 $Y_1(x_3)$  = UCT, etc.  
 $C_p(t)$  = kinetic acceleration or deceleration of the  $P$ -intensive  
 $K_{C_x^p}$  = acceleration or deceleration of the intensive as a result of introducing arrangement  $x$   
 $Y(x)$  = introduction of arrangement  $X$  carried out in the subsystem of macrosection  $AS$   
 $X$  = effect of change in arrangement  $X$  of subsystem  $AS$  on organic material or sewage at submicro level ( $AS_p, p'$ )  
 $a_{ij}$  = technological coefficient matrix  
 $T_p$  = nutrient load (transport)  
 $T_c$  = phosphorus load (transport)  
 $AS_c$  = dynamic effect of activated sludge technology in the settling pool.

### References

1. ANDREWS, I.—GRAEF: Dynamic modelling and simulation of the anaerobic digestion process *Advances in Chemistry Ser. 105*, 126 (1971)
2. ANDREWS, I.: Kinetics of biological processes used for wastewater treatment *Proc. Amer. Soc. Civ. Engr.*, 95, 95 (1970)
3. ANDREWS, I.: Dynamic models and computer simulation of wastewater systems Vanstelnkiste (ed.), Amsterdam p. 4557 (1975)
4. BECK, M. B.: Operation water quality management: Beyond planning and design IIASA, 32 (1981)
5. BELCOURT, A.: Effect of various anions and biological macromolecules on bacterial aggregation *J. Biol. Buccel.*, 3, 223 (1975)
6. BISOGNI, CAWERENCE: Relationship between the retention time of biological solids and settling characteristics of activated sludge *Water Research*, 5, 735 (1971)
7. BRIANT, J., WILCOX: Real-time simulation of the conventional activated sludge process *Proc. JACC*, p. 701 USA, California (1972)
8. BUSTY, ANDREWS, I.: Dynamic modelling and control strategies for the activated sludge process *Water Pollution Control Conference USA 1973*
9. BUZÁS, K.: Determination of the volume of reservoirs in unified and canalized systems *Hidrológia Közölny*, 8 (1978) (in Hungarian)
10. BUZÁS, K.: Pollutant load and its decreasing in receiving water bodies *Dissertation*, Budapest, Hungary (1983) (in Hungarian)
11. BUZÁS, K.: Technical-economical analysis of direct loads and their decreasing due to canalization of settlements IV. *Orsz. Vándorgyűlés, Győr, Hungary (1983)* (In Hungarian)
12. CURDS, C.: Theoretical study of factors influencing the microbial population dynamics *Water, Research*, 7, 1269 (1973)
13. CSAO, A.—KEINATH, T.: Influence of process loading intensity on sludge clarification and thickening characteristics *Water Research*, 13, 1213 (1979)
14. D'ASTOUS, F.: Analyzing environmental time series *J. Envir. Engr.*, 1979, 979
15. DILLON—RIGLER: Model predicting the phosphorus concentration in lake water *Res. Board. Can.*, 3, 1771 (1974)
16. DOBOLYI, E.: Data on the bottom sediment in Lake Balaton *Proc. 2nd Joint MTA, Vol. 2*, p. 66 (1980)

17. DOBOLYI, E.—BIDLÓ, G.: Some data on the bottom sediment of Lake Balaton *Hidrológiai Közöny.* 60, 72 (1980) (in Hungarian)
18. DULOVICS, D. et al.: Up-to-date canalization systems and the aspects of their design *Hidrológiai Közöny.* 6 (1978) (in Hungarian)
19. DULOVICS, D.: Dimensioning of artificial biological wastewater purification equipments *Tervezési Útmutató.* BME manuscript. (1980)
20. FRISK, T.: Modifications of phosphorus models *Aqua Fenicia.* 11, 7 (1981)
21. FÉNYES, L.: Thermostatics and thermodynamics *Műszaki Könyvkiadó, Budapest.* Hungary. 1967. (in Hungarian)
22. FOSTER, C.: Activated sludge surfaces in relation to the volume index of sludge *Water Research.* 5, 861 (1971)
23. FOSTER, C.: The nature of activated sludge flocs *Water Research.* 10, 25 (1976)
24. FORD—ECKENFELDER: Effect of process variables on sludge floc formation *JWPCF.* 39, 850 (1967)
25. GELENCSÉR, P.: Specific methods for the studying of water quality with respect to microelements and plant nutrients *VITUKI* 7783/3/294 (1980)
26. GRONA, A.: Kinetic parameters for municipal wastewater *JWPCF.* 51, 5 (1979)
27. GUSTAVSSON, I.: Parametric identification of multiple input Report 6907, Sweden, 1969
28. GUSTAVSSON, I.: Survey of applications of identification in chemical and physical processes Eybhoff (ed.) Amsterdam, Holland 1973.
29. HINO, M.: Ecohydrodynamics Adv. in Hydroscience Academic Press, Japan 1981, p. 12
30. KALLE MATTI: Phosphorus retention model *Water Research.* 60 (1981)
31. RECKLOW, K.: Uncertainty analysis, applied to Vollenweider's phosphorus loading criterion *JWPCF.* 51, 8 (1979)
32. KOHLÁR, GY. et al.: Eutrophisation processes in lakes *Acta Biol. Debrecenina* 13, 163 (1976)
33. LAPPALAINEN: Phosphorus load of lakes and a mathematical model for prognosis *Miljövard Descr. Bull.*, 1, 425 (1975)
34. LARSSON—SCHÖDER: Dynamic models for primary sedimentation in wastewater treatment Report-RE 146, Sweden, 1974
35. LAUBENBERGER, G.: Physical structure of activated sludge in aerobic stabilization *Water Research.* 5, 335 (1971)
36. LEENTVAAR—REBHUM: Strength of phenic hydroxide flocs *Water Research.* 17, 8, 895 (1983)
37. LITERÁTY, P. et al.: Investigation of the interstitial water in bottom sediments private communication
38. LOTKA, K.: Elements of mathematical biology Dover Publications N. USA, 1984.
39. MÁTÉ: The phosphorus retention on the sediment of Lake Balaton *Veszprém.* 1981
40. MÜLLER, P.: Sedimentbildung im Plattensee in Ungarn *Naturwissenschaften.* 56, 606 (1969)
41. OLÁH, J. et al.: Phosphorus metabolism of Lake Balaton *MTA Biol. Oszt. Közleményei.* 20, 11 (1973) (in Hungarian)
42. OLÁH, I.: Biomass and production of bacterioplankton in Lake Balaton *Hidrológiai Közöny.* 53, 181 (1973) (in Hungarian)
43. OLSSON, G. et al.: Control problems in wastewater treatment plants *Proc. Eng. Conf., USA California* 1973.
44. OLSSON, G.: State of the art in sewage treatment plant control. *Chemical Process Control.* USA California (1976)
45. ÖLLÖS, G.: The reuse of water in water supplies *Hidrológiai Közöny.* 1978, 12 (in Hungarian)
46. ÖLLÖS, G.: Reconstruction of water and canal works *Hidrológiai Közöny.* 1978, 3 (in Hungarian)
47. ROSERO, J. A.: Treatment of butcher's sewage BME Diplom work, Budapest 1982 (in Hungarian)
48. ROSERO, J. A.: Connection between environment protection and foreign trade *MKKE Szakdolgozat.* Budapest 1984
49. ROSERO, J. A.: Connection between wastewater treatment plant and receiving water body *Dissertation, Technical University, Budapest* (1985)
50. ROSERO, J. A.: Extensive-intensive connection between environment protection and foreign trade *Dissertation in economics.* MKKE Budapest 1986
51. SCHWARTZ, M.: Models for water reuse and wastewater planning *J. Environm. in.* 15, 109,5 (1984)
52. SCHAW, G., BROOKS, J.: Origin and development of living systems *Academic Press London* 1981.

53. SOMLYÓDI, L. et al.: Uncertain system identification and the prediction of water quality Pergamon Press, Oxford 1982.
54. SOMLYÓDI L.: Water quality management General report, case study 1982 Veszprém (in Hungarian)
55. SOMLYÓDI, L.: Eutrophisation of shallow lakes: modelling and management IIASA, CP-83-53 Laxenburg, Austria
56. SZŰCS, E.: Basis of similarity theory Műszaki Könyvkiadó, Budapest, 1967. (in Hungarian)
57. SZŰCS, E.: Dialogs about technical sciences Műszaki Könyvkiadó, Budapest, 1971. (In Hungarian)
58. SZŰCS, E.: Processes in sanitary engineering Műszaki Könyvkiadó, Budapest 1975. (In Hungarian)
59. SZŰCS, E.: Similitude and modelling Elsevier, Amsterdam 1980.
60. TÓTH, L.: Phosphorus metabolism of Lake Balaton Balaton Vízvédelmi Bizottság, 59 1975. (In Hungarian)
61. TÓTH, L.: Water quality of Lake Balaton No. 7 part of report VITUKI 7782/418 (1976) (In Hungarian)
62. TÓTH, L.: Sediment samples from Lake Balaton VITUKI 1781/3129 (1978) (in Hungarian)
63. VERHOFF, A.: Moment methods for the analysis of river models with application to point source phosphorus Water Research, 15, 493 (1980)
64. VOLLENVEIDER, R.: The scientific basis for lake and stream eutrophisation with particular reference to phosphorus and nitrogen OECD, Paris, DAS/CS/68.27.1. 1968.
65. VOLLENVEIDER, R.: Input-output models with special reference to the phosphorus loading concept in limnology Schweiz. Hydrol. 37, no. 1 53 (1975)
66. WELLS, C.: On the application of a non-linear regulator for a model of biological waste treatment IEEE. AC-16, 385 (1971)
67. WHITFIELD, P.: Selecting a method for estimating substance loading Water Resources Bull., 18, 2 (1982)
68. WIJ SENG: Phosphorus models for eutrophic lakes Water Research, 15, 493 (1976)
69. YOUNG, P.: The modelling and control of water quality Automatics, 10, 455 (1974)

Dr. José A. ROSERO }  
Dr. Kálmán BUZÁS } H-1521 Budapest